

NITROGEN AND PHOSPHOROUS REMOVAL OF PILOT-SCALE ANAEROBIC-ANOXIC-AEROBIC PROCESS UNDER PLATEAU ENVIRONMENTAL FACTORS

ZONG, Y. C.^{1,2,3,4,5,6} – HAO, K. Y.⁶ – LI, Y. W.⁶ – LU, G. H.^{1,2,3,4,5,6*} – HUANG, D. C.⁶

¹*Res. Institute of Tibet Plateau Ecology, Tibet Agriculture & Animal Husbandry University
Linzhi 860000, China*

²*Tibet Key Laboratory of Forest Ecology in Plateau Area, Ministry of Education
Linzhi 860000, China*

³*National Key Station of Field Scientific Observation & Experiment, Linzhi 860000, China*

⁴*Key Laboratory of Forest Ecology in Plateau Area, Tibet Autonomous Region
Linzhi 860000, China*

⁵*United Key Laboratories of Ecological Security, Tibet Autonomous Region
Linzhi 860000, China*

⁶*Water Conservancy Project & Civil Engineering College, Tibet Agriculture & Animal
Husbandry University, Linzhi 860000, China
(phone: +86-13062577435)*

**Corresponding author*

e-mail: ghlu@hhu.edu.cn; phone: +86-13062577435

(Received 3rd May 2019; accepted 11th Jul 2019)

Abstract. Considering the two unique factors of the plateau environment (i.e. water temperature and oxygen partial pressure), this paper explores the nitrogen and phosphorous removal of pilot-scale anaerobic-anoxic-aerobic (A2O) process. The experimental results show that none of the total phosphorous (TP), total nitrogen (TN) or NH₄⁺-N of the outlet water satisfied the Chinese national standard GB18918-2002, while COD only fulfilled the standard requirements under a few working parameters. The optimal values of the four working parameters were determined as: the optimal HRT=26.25 h, the optimal DO=3.0 mg/L, the optimal temperature=15°C. All these optimal values deviated greatly from the existing studies. The microbial densities corresponded poorly to the three working parameters, and were smaller than those under non-plateau environment; the number of indicator microorganisms had good correspondence with the optimal values of our experimental parameters.

Keywords: *anaerobic-anoxic-aerobic (A2O) process, water temperature, hydraulic retention time (HRT), dissolved oxygen (DO), nitrogen and phosphorous removal*

Introduction

The anaerobic-anoxic-aerobic (A2O) process, improved from the traditional activated-sludge process, has become the most popular sewage treatment method in Tibet (Chen et al., 2018), China. It is widely agreed that the effect of A2O is influenced by the unique environment factors of the plateau, namely, water temperature and pressure.

With a mean elevation of more than 4,000 m, Tibet can be categorized as a typical plateau region with low temperature. Taking Linzhi for instance, the domestic sewage falls in the range of 4°C and 14°C (Zong et al., 2018), which is a typical low-

temperature sewage. The water temperature in other parts of Tibet is theoretically below this range, because Linzhi is a city with below-average elevation in Tibet.

According to Lewis-Whitman's two-film theory (Ruiz-Urbieta et al., 1975), the maximal concentration of oxygen that can be dissolved in liquid medium decreases with the pressure. The atmospheric pressure is only 67.24 kPa in Linzhi. The elevation and atmospheric pressure of other cities in Tibet are listed in *Table 1* below.

Table 1. Elevation, pressure of Tibet in China

| | Lhasa | Changdu | Shigatse | Linzhi | Shannan | Naqu | Ali |
|-------------------------------|--------|---------|----------|--------|---------|--------|--------|
| Elevation (m) | 3658.0 | 3306.0 | 3836.0 | 3000.0 | 3551.7 | 4507.0 | 4278.0 |
| The local pressure (kPa) | 59.87 | 63.81 | 57.88 | 67.24 | 61.06 | 50.36 | 52.93 |
| Oxygen partial pressure (kPa) | 12.54 | 13.37 | 12.13 | 14.09 | 12.79 | 10.55 | 11.09 |
| Oxygen levels (%) | 59.87 | 63.81 | 57.88 | 67.24 | 61.06 | 50.36 | 52.93 |

Note: Standard Oxygen levels are 100%

Temperature is a main influencing factor of sewage treatment (Ai et al., 2014; Abourabia and Abdel Moneim, 2019). Its influence mainly exists in the following aspects: On the phosphorus removal of sludge denitrification, the release and absorption rates of phosphorus are changed under excessively high or low temperatures, and the proportion of denitrifying phosphorus accumulating organisms (DPAOs) in the activated sludge is greatly affected by temperature variation (Zhang et al., 2016; Wu, 2017); the nitrification capacity of the sewage treatment system is obviously weakened when the water temperature falls below 15°C, and basically disappears when the temperature drops below 4°C (He et al., 2010; Li et al., 2013). Thus, the nitrogen removal is severely inhibited under a low temperature (Li et al., 2014).

In view of the above, this paper explores the mechanism of nitrogen and phosphorus removal of A2O system, as a typical sewage treatment process. Starting from the microbial variation law, the author discussed the influence of water temperature, dissolved oxygen concentration, hydraulic retention time (HRT) and other factors on nitrogen and phosphorus removal, analyzed the operation features of A2O system at high elevations, and investigated the microbial features in the anaerobic section, anoxic section and aerobic section. The research discloses the mechanism of nitrogen and phosphorus removal in the reactor, shedding light on how plateau environmental factors affect the mechanism of nitrogen and phosphorus removal. Moreover, our research results lay a theoretical basis for the biological sewage treatment system in the plateau environment.

Materials and methods

Description of the A2O system and wastewater

As mentioned above, our experiment aims to disclose the law of removal rates under different water temperatures, HRTs and dissolved oxygens (DOs). For this purpose, the process flow of our experiment was designed as the following chart.

A pilot-scale A2O sewage treatment device was designed and fabricated with plexiglass. With an effective volume of 210 L, the device consists of 8 segments: the first 2 are anaerobic tanks, the middle 2 are anoxic tanks, and the last 4 are aerobic tanks.

The volume ratio of the anaerobic, anoxic and aerobic sections is 1:1:2. Besides, the effective volume of the sedimentation tank is 26.25 L. In both anaerobic and anoxic sections, each tank has a 50 rpm stirring device at the bottom; in each aerobic tank, there is an aerator for oxygen supply. Inflow, return sludge and nitrifying liquid are controlled by a peristaltic pump. To maintain a constant temperature, the water temperature was regulated by a constant temperature circulator. In each tank, a sampling hole was opened on the tank wall. Before the experiment, the activated sludge was cured for 32 d. The temperature, mixed liquor suspended solids (MLSS), and volume percent of MLSS after settling for 30 min (SV_{30}) were set to 22.5°C, 4,787 mg/L and 35%, respectively. The number of parallel samples per point is three, and the average of three samples is taken when the accuracy requirement is met, otherwise resampling is considered (*Figure 1*).

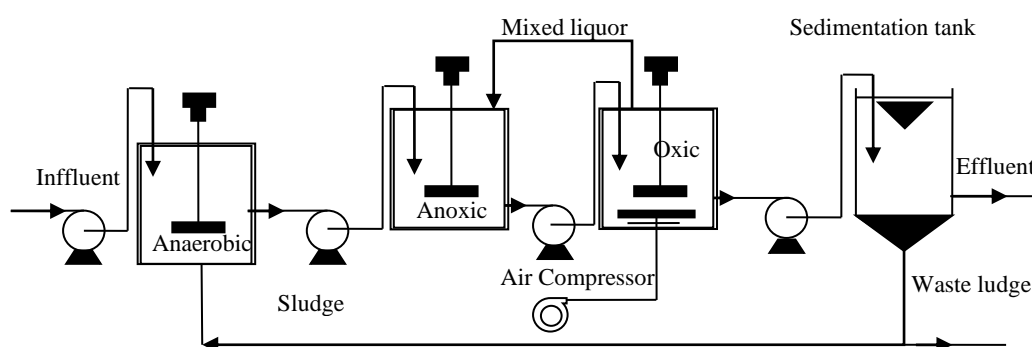


Figure 1. Schematic of anaerobic-anoxic-aerobic (A2O) process

The urban domestic sewage in Linzhi was directly adopted for our experiments. The main water quality indices of the sewage are given in the table below (*Table 2*).

Table 2. Quality indicators of sewage

| pH | DO (mg/L) | Temperature (°C) | COD (mg/L) | TN (mg/L) | TP (mg/L) | NH ₄ ⁺ -N (mg/L) |
|-----------|-----------|------------------|---------------|--------------|------------|--|
| 6.99~8.41 | 1.12~5.08 | 8.98~29.10 | 149.48~526.36 | 36.51~147.08 | 3.96~11.56 | 30.81~54.16 |

Operation of the A2O device

The operation of the A2O device was studied under three working parameters, i.e. water temperature, hydraulic retention time (HRT) and DO, in order to disclose the law of removal rates under different water temperatures, HRTs and DOs. The control plans for the three parameters are specified below.

On water temperature control, the inlet water flow was designed as 10.0±0.1 L/s, HRT as 21.0±0.2 h (the HRT ratio between anaerobic tank, anoxic tank and aerobic tank=1: 1: 2), DO as 2.0±0.1 mg/L, the reflux ratio of the mixed liquor $R_i=200\%$, and the reflux ratio of the sludge $R=100\%$. Both the mixed liquor and the sludge were continuously refluxed. The temperature was changed by an electric heater between five levels: 10, 15, 20, 25 and 30°C. The water samples were collected 72 h after the temperature reached the design temperature. The temperature was controlled with an error of or less than 0.1°C.

On HRT control, the inlet water temperature was designed as $20\pm 0.1^{\circ}\text{C}$, the HRT ratio between anaerobic tank, anoxic tank and aerobic tank as 1: 1: 2, DO as 2.0 ± 0.1 mg/L, the reflux ratio of the mixed liquor $R_i=200\%$, and the reflux ratio of the sludge $R=100\%$. Both the mixed liquor and the sludge were continuously refluxed. HRT was adjusted by changing the inlet water flow between five levels: 4, 8, 12, 16 and 20 L/s. HRTs corresponding to the five levels were respectively 52.5, 26.26, 17.5, 13.125 and 10.5 h. The water samples were collected 72 h after the change of the inlet water flow. The flow was controlled with an error of or less than ± 0.1 L/s.

On DO control, the inlet water temperature was designed as $20\pm 0.1^{\circ}\text{C}$, the inlet water flow as 10.0 ± 0.1 L/s, HRT as 21.0 ± 0.2 h (HRT ratio between anaerobic tank, anoxic tank and aerobic tank=1:1:2), the reflux ratio of the mixed liquor $R_i=200\%$, and the reflux ratio of the sludge $R=100\%$. Both the mixed liquor and the sludge were continuously refluxed. DO was altered between 10 levels (1, 2, 2.5, 2.8, 3, 3.2, 3.5, 4, 4.5 and 5 mg/L) by changing the amount of blast aeration. The water samples were collected 72 h after DO reached the design level. DO was controlled with an error of or less than ± 0.1 mg/L.

Analytical methods

According to the working standards, the experimental indices were respectively measured by the following methods: COD was determined by the potassium dichromate method, the TN by the ion chromatography, TP by potassium persulfate oxidation, MLSS by the gravimetric method, the SV_{30} by the standard method, DO by membrane electrode method, pH by the portable pH meter, $\text{NH}_4^+\text{-N}$ by the Nessler's reagent photometry, the water temperature by water thermometer, the inlet water flow by float flowmeter, and the microorganisms by biochemical microscope counting.

Results

Sewage treatment at different HRTs

Table 3 displays the operation parameters like DO, MLSS, pH and temperature at different HRTs.

Table 3. Process Parameter at different hydraulic retention times (HRTs)

| Designed Flow (L/h) | 4 | 8 | 12 | 16 | 20 |
|------------------------------------|-------|-------|-------|-------|-------|
| HRTs (h) | 52.50 | 26.25 | 17.50 | 13.13 | 10.50 |
| DO (mg/L) | 3.16 | 2.30 | 1.64 | 1.98 | 2.04 |
| pH | 8.10 | 8.00 | 8.27 | 8.40 | 8.41 |
| MLSS (mg/L) | 964 | 996 | 1084 | 930 | 920 |
| SV_{30} (%) | 10 | 10 | 10 | 10 | 10 |
| Temperature ($^{\circ}\text{C}$) | 21.4 | 20.0 | 19.5 | 20.7 | 20.6 |

Effect on removal rates

Figure 2 provides the removal rates of COD, TP, TN and $\text{NH}_4^+\text{-N}$ in anaerobic, anoxic and aerobic tanks at different HRTs.

As shown in *Figure 2(a)*, the COD removal rate ranged between 69.65% and 93.98% under the five HRTs; in descending order of the COD removal rate, the inlet water flows were ranked as 8, 12, 16, 4 and 20 L/h, and HRTs as 26.25, 17.50, 13.13, 52.50 and 10.50 h; the optimal HRT was much greater than 7~14 h; the highest COD removal rate was observed at the design inlet water flow of 8 L/h; the removal effect was obvious in anaerobic and aerobic sections, but the removal rate was not significantly enhanced in the anoxic section.

As shown in *Figure 2(b)*, the TP removal rates fell in the range of 4.53%~64.56% under the five HRTs; in descending order of TP removal rate, the inlet water flows were ranked as 8, 12, 20, 16 and 4 L/h, and HRTs as 26.25, 17.50, 10.50, 13.13 and 52.50 h; the optimal HRT was much greater than 7~14 h; the highest TP removal rate was observed at the design inlet water flow of 8 L/h; the removal effect was obvious in anaerobic and aerobic sections, but the removal rate was not significantly enhanced in the anoxic section.

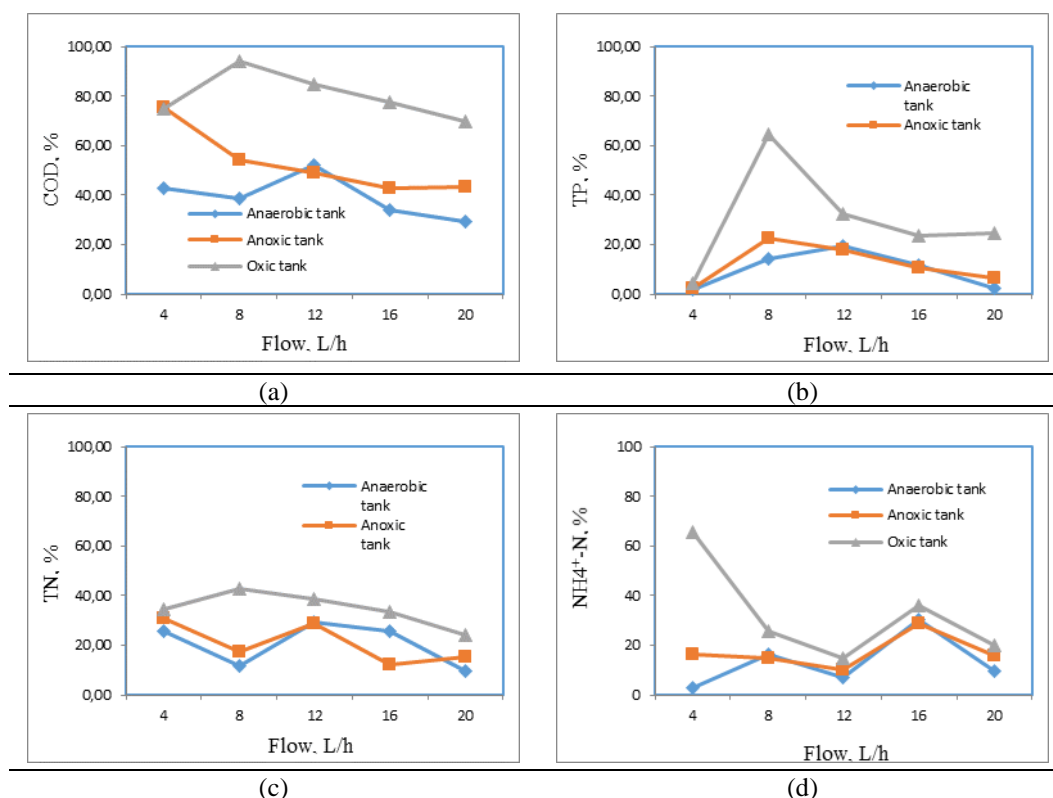


Figure 2. Removal rate at different hydraulic retention times (HRTs)

As shown in *Figure 2(c)*, the TN removal rates shifted from 23.92% to 42.85% under the five HRTs; in descending order of the TN removal rate, the inlet water flows were ranked as 8, 12, 4, 16 and 20 L/h, and HRTs as 26.25, 17.50, 52.50, 13.13 and 10.50 h; the optimal HRT was much greater than 7~14 h; the highest TN removal rate was observed at the design inlet water flow of 8 L/h; the removal effect was obvious in anaerobic and aerobic sections, but the removal rate was not significantly enhanced in the anoxic section.

As shown in *Figure 2(d)*, the $\text{NH}_4^+\text{-N}$ removal rates changed within 14.92% and 65.47%; in descending order of the $\text{NH}_4^+\text{-N}$ removal rate, the inlet water flows were ranked as 4, 16, 8, 20, 12 L/h, and the HRTs as 52.50, 13.13, 26.25, 10.50 and 17.50 h; the optimal HRT was much greater than 7-14 h; the highest $\text{NH}_4^+\text{-N}$ removal rate was observed at the design inlet water flow of 4 L/h; the removal effect was obvious in the aerobic sections, but the removal rate was not significantly enhanced in the anaerobic or anoxic section.

To sum up, the optimal HRT was 26.25 h among the five designed HRTs; HRT increased more significantly in the plateau environment than other regions.

Microbial response

The microbial density and indicator microorganisms under the above five HRTs are shown in *Figure 3*.

It can be seen from *Figure 3(a)* that, under the five HRTs, the different levels of inlet water flows could be ranked as 20, 16, 8, 4 and 12 L/h in descending order of the number of microorganisms in the anaerobic section, 4, 16, 12, 20 and 8 L/h in the anoxic section, and 20, 4, 12, 16 and 8 L/h in the aerobic section. Since the microbial effect is correlated with microbial density, microbial residence time, and inlet/outlet water quality, it is not reasonable to evaluate the microbial response with the bulk density of the microorganisms.

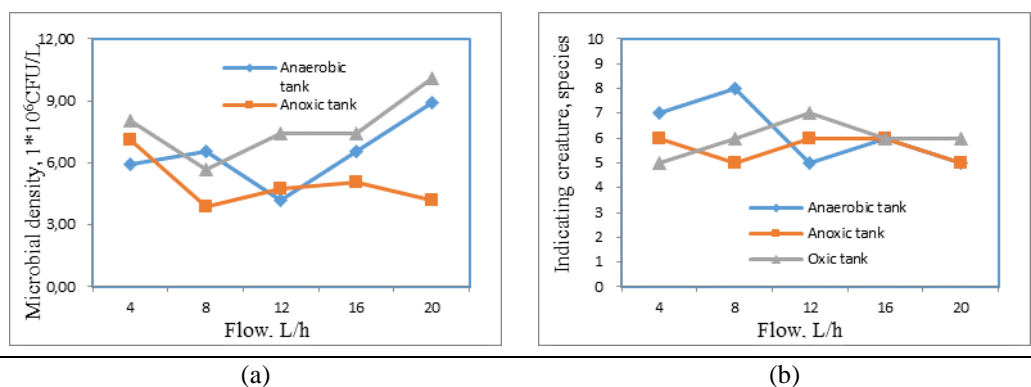


Figure 3. Microbial response at different hydraulic retention times (HRTs)

It can be seen from *Figure 3(b)* that, under the five HRTs, the different levels of inlet water flows could be ranked as 8, 4, 16, 12 and 20 L/h in descending order of the diversity of indicator microorganisms in the anaerobic section, 16, 4, 12, 8 and 20 L/h in the anoxic section, and 12, 8, 16, 20 and 4 L/h in the aerobic section. The indicator microorganisms reached the peak diversity at 8 L/h. The most populous microorganisms include vorticella, trochilia, rotifera and oxytricha.

The microbial analysis show that the microbial density is not directly correlated with the optimal HRT, because the treatment effect depends on retention time and inlet/outlet water quality, in addition to microbial density; moreover, the indicator microorganism analysis confirmed that the indicator microorganisms reached the peak diversity at 8 L/h and the most populous microorganisms were vorticella, trochilia, rotifera and oxytricha.

Sewage treatment at different DOs

Table 4 lists the operation parameters like DO, MLSS, pH and temperature at different DOs.

Table 4. Process parameter at different dissolved oxygens (DOs)

| Designed DO (mg/L) | 1.0 | 2.0 | 2.5 | 2.8 | 3.0 | 3.2 | 3.5 | 4.0 | 4.5 | 5.0 |
|----------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| HRT (h) | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 |
| DO (mg/L) | 1.12 | 2.01 | 2.42 | 2.82 | 2.92 | 3.20 | 3.54 | 3.90 | 4.50 | 5.08 |
| pH | 7.43 | 7.21 | 7.55 | 8.18 | 7.61 | 7.91 | 8.03 | 7.46 | 7.43 | 7.34 |
| MLSS (mg/L) | 247 | 568 | 714 | 928 | 1089 | 1043 | 1015 | 957 | 724 | 616 |
| SV ₃₀ (%) | 8 | 9 | 10 | 10 | 10 | 10 | 10 | 10 | 9 | 8 |
| Temperature (°C) | 20.10 | 21.00 | 18.80 | 20.00 | 18.70 | 20.00 | 20.40 | 19.70 | 19.80 | 20.80 |

Effect on removal rates

Figure 4 gives the removal rates of COD, TP, TN and NH₄⁺-N in anaerobic, anoxic and aerobic tanks at different DOs.

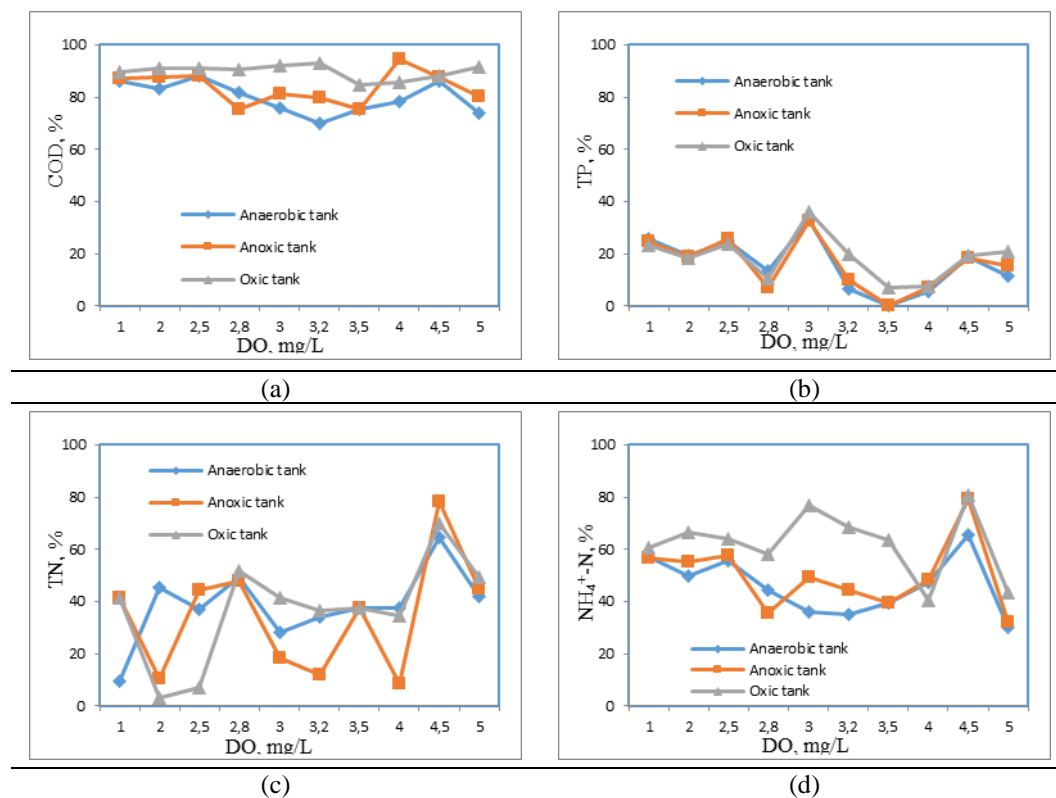


Figure 4. Removal rate at different dissolved oxygens (DOs)

As shown in Figure 4(a), the COD removal rate ranged between 84.59% and 92.84% under the ten DOs; in descending order of the COD removal rate, the DOs were ranked as 3.2, 3.0, 5.0, 2.5, 2.0, 2.8, 1.0, 4.5, 4.0 and 3.5 mg/L; the optimal DO was far higher

than 2.0 mg/L, but the COD differed slightly between the working parameters; the COD removal rate peaked at the design DO of 3.2 mg/L; the removal effect was obvious in anaerobic section, but the removal rate was not significantly enhanced in the anoxic or aerobic section.

As shown in *Figure 4(b)*, the TP removal rate fluctuated between 7.27% and 35.84% under the ten DOs; in descending order of the TP removal rate, the DOs were ranked as 3, 2, 1, 5, 3.2, 4.5, 2, 2.8, 4 and 3.5 mg/L; the optimal DO was far higher than 2.0 mg/L; the highest TP removal rate appeared at the design DO of 3 mg/L; the removal effect was obvious in anaerobic section, but the removal rate was not increased in the anoxic or aerobic section.

As shown in *Figure 4(c)*, the TN removal rate fell between 7.27% and 35.84% under the ten DOs; in descending order of the TN removal rate, the DOs were ranked as 3, 2, 1, 5, 3.2, 4.5, 2, 2.8, 4 and 3.5 mg/L; the optimal DO was much higher than 2.0 mg/L; the highest TN removal rate appeared at the design DO of 3 mg/L; the removal effect was obvious in aerobic section, but the removal rate was not increased significantly in the anoxic or anaerobic section.

As shown in *Figure 4(d)*, the NH_4^+ -N removal rate changed between 80.76% and 40.35% under the ten DOs; in descending order of the NH_4^+ -N removal rate, the DOs were ranked as 4.5, 3.0, 3.2, 2.0, 2.5, 3.5, 1.0, 2.8, 5.0 and 4.0 mg/L; the optimal DO was much higher than 2.0 mg/L; the NH_4^+ -N removal rate reached the maximum at the design DO of 4.5 mg/L; the removal effect was obvious in aerobic section, but the removal rate was not increased significantly in the anoxic or anaerobic section.

In summary, the optimal DO was 3 mg/L among the ten designed DOs; the DO in the plateau environment was much greater than that in the other regions; however, the removal rates of TP, TN and NH_4^+ -N were not high under plateau environmental factors; in particular, only about 35% of TP were removed.

Microbial response

The microbial density and indicator microorganisms under the above seven DOs are displayed in *Figure 5*.

It can be seen from *Figure 5(a)* that the seven DOs could be ranked as 2.8, 4.5, 2.5, 5, 2, 3.2 and 1 mg/L in descending order of the number of microorganisms in the anaerobic section, 2.5, 5.0, 2.8, 3.2, 2, 4.5 and 1 mg/L in the anoxic section, and 2.5, 2.8, 5.0, 3.2, 4.5, 2 and 1 mg/L in the aerobic section. Since the microbial effect is correlated with microbial density, microbial residence time, and inlet/outlet water quality, it is not reasonable to evaluate the microbial response with the bulk density of the microorganisms.

It can be seen from *Figure 5(b)* that the seven DOs could be ranked as 2.8, 4.5, 1, 2, 2.5, 5 and 3.2 mg/L in descending order of the diversity of indicator microorganisms in the anaerobic section, 1, 5, 2.5, 3.2, 4.5, 2.8 and 2 mg/L in the anoxic section, and 5, 2, 2.5, 3.2, 1, 2.8 and 4.5 mg/L in the aerobic section. The indicator microorganisms reached the peak diversity at 5 mg/L, when the most populous microorganisms were vorticella and rotifera. Meanwhile, the most populous microorganisms became vorticella, rotifera and trochilia at the DO of 3 mg/L.

The microbial analysis show that the microbial density is not directly correlated with the optimal DO, because the treatment effect depends on retention time and inlet/outlet water quality, in addition to microbial density; moreover, the indicator microorganism

analysis confirmed that the indicator microorganisms reached the peak diversity at 5mg/L, when the most populous microorganisms were vorticella and rotifera.

Sewage treatment at different water temperatures

Table 5 lists the operation parameters like DO, MLSS, pH and HRT at different water temperatures.

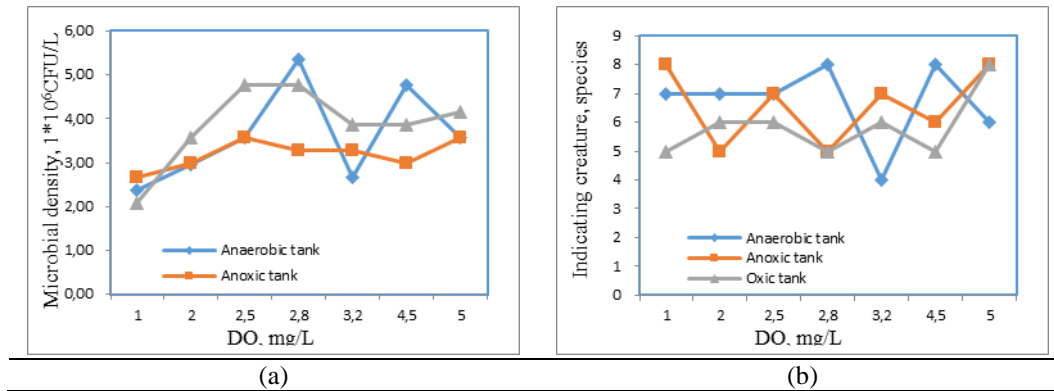


Figure 5. Microbial response at different dissolved oxygens (DOs)

Table 5. Process parameter at different temperatures

| Designed Temperature(°C) | 10 | 15 | 20 | 25 | 30 |
|--------------------------|-------|-------|-------|-------|-------|
| HRT(h) | 26.25 | 26.25 | 26.25 | 26.25 | 26.25 |
| DO(mg/L) | 1.80 | 1.87 | 1.76 | 1.80 | 1.63 |
| pH | 7.52 | 7.69 | 7.58 | 7.52 | 8.18 |
| MLSS(mg/L) | 1137 | 2132 | 1889 | 1137 | 983 |
| SV30(%) | 15 | 17 | 15 | 15 | 13 |
| Temperature(°C) | 8.98 | 14.8 | 17.5 | 22.4 | 29.1 |

Effect on removal rates

Figure 6 shows the removal rates of COD, TP, TN and NH₄⁺-N in anaerobic, anoxic and aerobic tanks at different water temperatures.

As shown in Figure 6(a), the COD removal rate ranged between 70.06% and 84.33% under the five water temperatures; in descending order of the COD removal rate, the water temperatures were ranked as 15, 30, 25, 10 and 20°C; the optimal water temperature was obviously 15°C, as the COD removal rate reached the peak value under this temperature; the removal effect was obvious in anaerobic section, but the removal rate was not significantly enhanced in the anoxic or aerobic section.

As shown in Figure 6(b), the TP removal rate fluctuated between 3.54% and 75.55% under the five water temperatures; in descending order of the TP removal rate, the water temperatures were ranked as 15, 20, 30, 10 and 25°C; the optimal water temperature was obviously 15°C, as the TP removal rate was the highest under this temperature; the removal effects were obvious in anaerobic and aerobic sections, but the removal rate was not significantly enhanced in the anoxic section.

As shown in *Figure 6(c)*, the TN removal rate fell between 36.60% and 77.98% under the ten DOs; in descending order of the TN removal rate, the water temperatures were ranked as 10, 25, 15, 30 and 20°C; the optimal water temperatures were 10 and 25°C, as the peak TN removal rate was measured under the two temperatures; the removal effects were obvious in anaerobic and aerobic sections, but the removal rate was not significantly enhanced in the anoxic section.

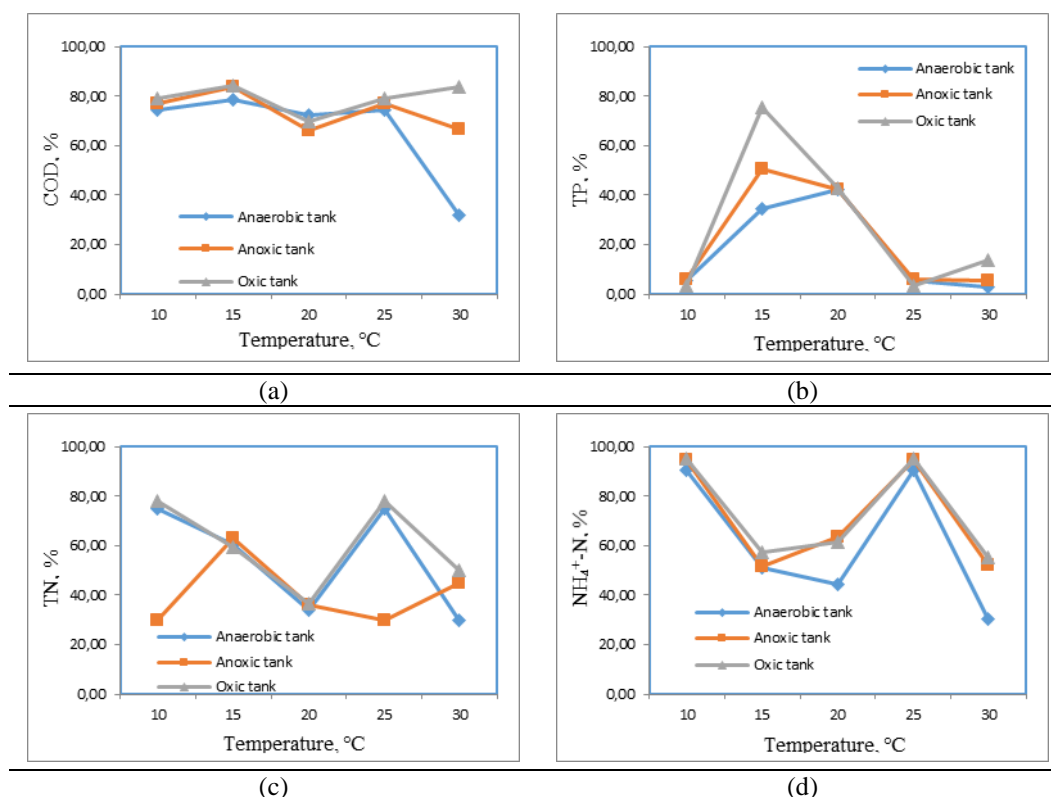


Figure 6. Removal rate at different temperatures

As shown in *Figure 6(d)*, the $\text{NH}_4^+\text{-N}$ removal rate changed between 55.44% and 95.44% under the five water temperatures; in descending order of the $\text{NH}_4^+\text{-N}$ removal rate, the water temperatures were ranked as 10, 25, 20, 15 and 30°C; the optimal temperatures were 10 and 25°C, as the peak $\text{NH}_4^+\text{-N}$ removal rate was obtained under these two temperatures; the removal effect was obvious in anaerobic section, but the removal rate was not significantly enhanced in the anoxic or aerobic section.

Overall, the optimal water temperature was 15°C among the five designed water temperatures; the optimal temperature in the plateau environment was lower than that in the other regions.

Microbial response

The microbial density and indicator microorganisms under the above five water temperatures are displayed in *Figure 7*.

It can be seen from *Figure 7(a)* that the five water temperatures could be ranked as 20, 15, 10, 25 and 30°C in descending order of the number of microorganisms in the anaerobic section, 30, 25, 10, 20 and 15°C in the anoxic section, and 15, 30, 25, 10 and

20°C in the aerobic section. Since the microbial effect is correlated with microbial density, microbial residence time, and inlet/outlet water quality, it is not reasonable to evaluate the microbial response with the bulk density of the microorganisms.

It can be seen from *Figure 7(b)* that the five water temperatures could be ranked as 20, 10, 25, 15 and 30°C in descending order of the diversity of indicator microorganisms in the anaerobic section, 20, 10, 25, 15 and 30°C in the anoxic section, and 15, 30, 10, 25 and 20°C in the aerobic section. The indicator microorganisms reached the peak diversity at the water temperature of 20°C, when the most populous microorganisms were vorticella, trochilia and oxytricha. Meanwhile, the most populous microorganisms were also vorticella, trofitera and trochilia at the water temperature of 15°C.

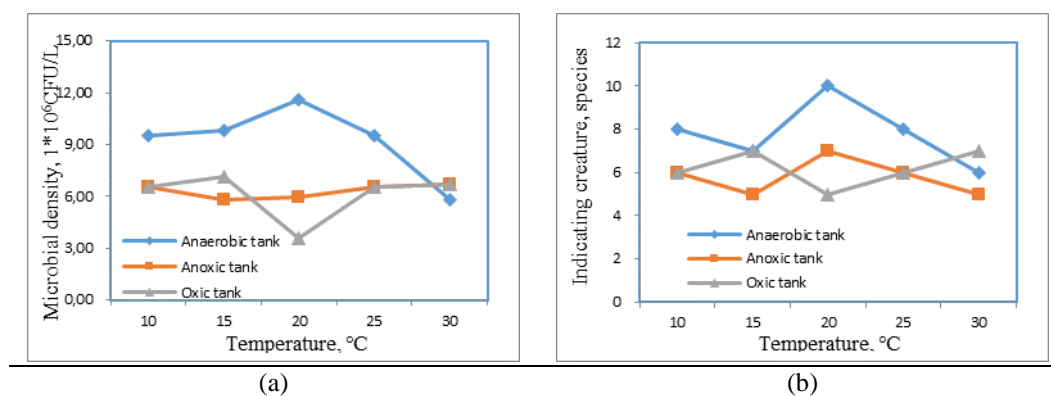


Figure 7. Microbial response at different temperatures

The microbial analysis show that the microbial density is not directly correlated with the optimal water temperature, because the treatment effect depends on retention time and inlet/outlet water quality, in addition to microbial density; moreover, the indicator microorganism analysis found that the indicator microorganisms reached the peak diversity at 15°C, when the most populous microorganisms were vorticella, rotifera and oxytricha; furthermore, it should be noted that the diversity did not increase with the water temperature (Tian et al., 2013).

Discussion

Discussion of HRTs

Some of the recent studies on the optimal HRT are as follows. Wang et al. (2014) suggested that the optimal HRT was 8~12 h for the processing of high-load sewage by A2O and electro-coagulation. Ye et al. (2018) put the optimal HRT to 6 h for the removal of organic matters and nutrients from urban sewage with dual-A2O (D-A2O).

In our experiment, the optimal HRT is determined as 26.25 h based on the removal rates and microorganism response. This conclusion differs greatly from the results of the previous studies. The difference may be attributed to the environmental factor of low water temperature. The microbial density in low-temperature water body treated by the A2O is generally considered as 10⁸CFU/L (Yang, 2017). In our experiment, however, the microbial density was merely 10⁶CFU/L under the plateau environmental factors.

Discussion of DO

Large numbers of DOs come to the anoxic unit through International recirculation flow and destroyed the hypoxic environment, affecting the nitrogen removal (Li et al., 2012). With respect to total nitrogen removal, nitrification–denitrification at low DO levels of 0.3–0.5 mg/L was essentially equal to the complete nitrification–denitrification at DO levels of 1.5–2.5 mg/L with the addition of external carbon sources (Zeng et al., 2010). Under the conditions of 11 h HRT, 1.0–2.0 mg/L DO concentration, 200% mixture reflux proportion, 80% sludge reflux proportion and 20d sludge age, the effluent concentration can achieve the first order A standard of Discharge Standard of Pollutants for Municipal Wastewater Treatment Plant (Zhang et al., 2011). Controlling to DO at 0.8–1.5 mg/L, the treatment efficiency of full-scale Biolac/A²O process was near optimal with the total nitrogen efficiency of 69.45% (Ju et al., 2013). DO levels in the range of 0.5 to 3.5 mg/L in the aeration basin did not have significant impact on effluent ortho-P concentration in a completely mixed basin within the EBPR process (Gu et al., 2006). In our experiment, the optimal DO is determined as 3.0 mg/L based on the removal rates and microorganism response. This conclusion deviates greatly from the results of the previous studies. A possible cause for the deviation lies in the unique environmental factors of the plateau.

Discussion of temperature

The temperature and functions of A2O process were closely related (Zhao et al., 2010). Under the condition of average temperature of 14.2°C and carbon-to-nitrogen ratio of 4.81, enhanced nitrogen and phosphorus removal was achieved (Wang, 2010). The changes of temperature could affect the community structure and the Shannon diversity index of nitrifying bacteria (Tao et al., 2009). In our experiment, the optimal water temperature was determined as 15°C based on the removal rates and microorganism response. This conclusion was very inconsistent with the results of the previous studies, which may be ascribed to the unique properties of microorganisms at low temperatures on the plateau. Because Nitrification rates at high altitude aquatic ecosystems are scarce (Hayden and Beman, 2014; Molina et al., 2018).

Conclusions

Under the unique environmental factors of plateau, this paper designs a pilot-scale A2O device to explore the effects of DO, HRT and water temperature on the removal rates of COD, TP, TN and NH₄⁺-N, and the law of microbial response to these working parameters. The results show that none of the TP, TN or NH₄⁺-N of the outlet water satisfied the Chinese national standard GB18918-2002, while the COD only fulfilled the standard requirements under a few working parameters. The optimal values of the four working parameters were determined as: the optimal HRT=26.25 h, the optimal DO=3.0 mg/L, the optimal temperature=15°C. All these optimal values deviated greatly from the existing studies. The microbial densities corresponded poorly to the three working parameters, and differed significantly from those under normal conditions; the number of indicator microorganisms was basically consistent with the optimal values of our experimental parameters. In addition, the SV30 (Xu et al., 2013) and NLSS (Zhang et al., 2018) were both low in our experiment.

Acknowledgements. This work was supported by the National Natural Science Foundation of China (NO.51868069, 51769034), Natural Science Foundation of Tibet (NO. XZ 2018 ZR G-20), the Program for Scientific Research Innovation Team in Colleges and Universities of Tibet Autonomous Region, Snowy Plateaus of Tibet Agriculture and Animal Husbandry College (Study on the operation status of typical sewage treatment plants in Tibet in China).

REFERENCES

- [1] Abourabia, A. M., Abdel Moneim, S. A. (2019): Analytical solution of sea water steady magneto-hydrodynamic equations subjected to stretching sheet under induced magnetic field and heat transfer. – *Mathematical Modelling of Engineering Problems* 6(1): 141-151.
- [2] Ai, S. S., Zhang, X. H., Xiao, Y. B. (2014): Study on characteristics of activated sludge at low temperature. – *Applied Mechanics & Materials* 675-677: 574-577.
- [3] Chen, X. Y., Hao, Ka. Y., Su, D. (2018): Characteristic study on wastewater treatment in high altitude area by A2/O process. – *Technology of Water Treatment* 38(6): 93-96.
- [4] Gu, A. Z., Hughes, T., Fisher, D. (2006): The devil is in the details: full-scale optimization of the EBPR process at the city of las vegas WPCF. – *Proceedings of the Water Environment Federation* 2006(7): 5110-5130.
- [5] Hayden, C. J., Beman, M. (2014): High abundances of potentially active ammonia-oxidizing bacteria and archaea in oligotrophic, high-altitude lakes of the sierra nevada, California, USA. – *Plos One* 9(11): e111560.
- [6] He, J. G., Ke, L., Han, B. P. (2010): Study on the operational characteristics of hybrid A2/O process at low temperature. – *Applied Mechanics & Materials* 39: 326-331.
- [7] Ju, Y. K., Wang, H. L., Zhang, Q. (2013): Effect of dissolved oxygen on nitrogen and phosphorus removal rate in biolac process. – *Advanced Materials Research* 779-780: 1629-1633.
- [8] Li, Y. F., Yang, J. Y., Zhang, G. C. (2012): Effects of Aeration on Nitrogen and Phosphate Removal with A2O Process. – *Advanced Materials Research* 622-623: 1738-1741.
- [9] Li, Y. F., Yang, J. Y., Zhang, G. C. (2013): Effects of aeration on nitrogen and phosphate removal with A2O process. – *Advanced Materials Research* 622-623: 1738-1741.
- [10] Li, S. M., Du, G. S., Tang, F. B. (2013): Nitrogen and phosphorus removal of modified A2/O process on low-carbon domestic sewage under low temperature. – *Advanced Materials Research* 777: 187-191.
- [11] Li, S. M., Hao, T., Wang, R. B. (2014): Operation of modified A~2/O process at low temperature and different sludge loadings. – *China Water & Wastewater* 2014(13): 64-68.
- [12] Molina, V., Dorador, C., Fernández, C. (2018): The activity of nitrifying microorganisms in a high altitude Andean wetland. – *FEMS Microbiology Ecology* 94(6).
- [13] Ruiz-Urbieta, M., Sparrow, E. M., Parikh, P. D. (1975): Two-film reflection polarizers: theory and application. – *Applied Optics* 14(2): 486-492.
- [14] Tao, F., Huang, Y., Gao, S., Huang, M. S., Chen, C. (2009): Application of PCR-DGGE to analyze the effect of temperature on structure of nitrifying bacteria in A/O system (Chinese). – *Journal of East China Normal University*.
- [15] Tian, X., Ai, S. S., Zuo, Y. (2013): Study on activated sludge microorganisms of northern winter sewage treatment plant. – *Applied Mechanics and Materials* 361-363: 1032-1035.
- [16] Wang, J. H. (2010): Biological nutrients removal from domestic wastewater with low carbon-to-nitrogen ratio in A~2O-BAF system at low temperature. – *China Environmental Science* 30(9): 1195-1200.
- [17] Wang, W., Chen, S., Bao, K. (2014): Enhanced removal of contaminant using the biological film, anoxic-anaerobic-aerobic and electro-coagulation process applied to high-load sewage treatment. – *Environmental Technology* 35(7): 833-840.

- [18] Wu, C., Guo, L. (2017): Influence of temperature and dissolved oxygen on nitrogen and phosphorus removal of integrated bioreactor. – *International Journal Bioautomation* 21(1): 207-216.
- [19] Xu, X. P., Tao, X. W., Du, J., Wu, F. S. (2013): Effect of SRT on nitrogen and phosphorus removal in A2/O process. – *China Water & Wastewater* 29(21): 69-71.
- [20] Yang, T. (2017): Study on degradation Regularity of three characteristic pollutants of pharmaceutical park tail water in a2o treatment process. – Liaoning University.
- [21] Ye, C., Zhou, Z., Li, M. (2018): Evaluation of simultaneous organic matters and nutrients removal from municipal wastewater using a novel bioreactor (D-A2O) system. – *Journal of Environmental Management* 218: 509-515.
- [22] Zeng, W., Li, L., Yang, Y. (2010): Nitritation and denitritation of domestic wastewater using a continuous anaerobic–anoxic–aerobic (A2O) process at ambient temperatures. – *Bioresource Technology* 101(21): 8074-8082.
- [23] Zhang, S. R., Zhang, T. J., Liu, J. L. (2011): Study on A2O Method for co-treatment of landfill leachate and municipal sewage. – *Advanced Materials Research* 356-360: 2908-2913.
- [24] Zhang, L., Zhuang, Y., Wang, X., Zhang, H. (2016): Effect of temperature on denitrifying phosphorus removal efficiency using modified A2/O process. – *Transactions of the Chinese Society of Agricultural Engineering*.
- [25] Zhang, J. X., Sun, W. G., Niu, F. S., Wang, L., Zhao, Y. W., Han, M. M. (2018): Atmospheric sulfuric acid leaching thermodynamics from metallurgical zinc-bearing dust sludge. – *International Journal of Heat and Technology* 36(1): 229-236.
- [26] Zhao, F., Dai, X. C., Huang, M. S. (2010): Influence of temperature on nitrogen removal in A2/O process. – *Environmental Science & Technology* 33(3): 49-53.
- [27] Zong, Y. C., Zhang, Y. H., Lu, G. H. (2018): Study on Process Characteristics of High Altitude A2/O Process Based on Principal Component Analysis. – *Technology of water treatment* 38(9): 116-119.