

GENETIC DIVERSITY AMONG ENDANGERED FRESHWATER MUSSELS BASED ON SHELL DNA SEQUENCES

DOĞANKAYA, L.* – COŞKUN, T. – GÜLTEKİN, T.

Ankara University, Faculty of Agriculture, Department of Fisheries and Aquaculture
Engineering, 06110 Dışkapı, Ankara, Turkey
(phone: +90-312-596-1438; fax: +90-312-318-5298)

*Corresponding author

e-mail: dogankaya@ankara.edu.tr; phone: +90-312-596-1439; fax: +90-312-318-5298

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Abstract. Molecular genetic information on mollusk species is being increasingly considered for identification and taxonomic status. In this study, we examined the potential of shell material of freshwater mussels from four different locations in Turkey for DNA sequence analysis. A total number of 212 specimens including empty shells were sampled from a river and three lakes. Mitochondrial DNA was extracted from ground shell material with isolation kits and tissue material was also used for reference. Even though the majority of empty shells did not produce high-quality material, DNA extraction from fresh shells had similar results with tissue material. Mitochondrial 28S rDNA partial sequences were analyzed using the neighbour-joining method for genetic divergence. Phylogenetic analysis revealed two well-differentiated clades of *Anodonta cygnea* and *Unio crassus*. The results clearly indicated that 28S rRNA sequences obtained from shell material are informative enough to reveal major clades, haplotypes and phylogenetic relationship between four populations. These findings support that quite fresh shell material from freshwater mussels can be a convenient DNA source for molecular studies.

Keywords: Unionidae, phylogeny, shell DNA, bivalve diversity, 28S rDNA

Introduction

Freshwater mussels as prehistoric bivalve inhabitants of inland waters are represented on all continents except Antarctica. They are classified into six families, about 180 genera and 800 species (Bogan, 2008). Bivalves are remarkable aquatic organisms with heavy impact on ecosystems through filtration, bio-accumulation, energy transfer, being host for other organisms and their role in food web. Such factors as pollution, loss of host fishes and the introduction of non-native species are causing changes in diversity and population structure (Lopez-Lima et al., 2017). Freshwater mussels are globally in decline and are among the most endangered aquatic species (Strayer, 2008). Lopez-Lima et al. (2017) recorded 224 (44%) of the 511 freshwater mussel species as Near Threatened or Threatened in the 2015 *IUCN Red List of Threatened Species*. Despite having highly important ecosystem roles, gaps exist on biology and population structure of freshwater bivalves still standing as the primary obstacle to management and conservation.

Recently, European freshwater mussels have also been gaining increasing attention and 16 species have been recognized in Unionida Margaritiferidae: 2 and Unionidae: 14 (Lopez-Lima et al., 2017). The current taxonomy of freshwater mussels is based on morphological characteristics and studies on genetic diversity are very scarce. Due to variable characteristics as size and shell shape which are affected by the current velocity, substratum and environment, freshwater mussels exhibit high phenotypic plasticity (Baker et al., 2004).

Population size, distribution and species data are primary requirements for successful conservation management so the taxonomic status of the order Unionida is to be resolved and DNA sequence would serve valuable data for systematic updates.

Mantle clipping is the widespread method for tissue sampling from mollusks in genetic studies but non-invasive sampling methods are needed for endangered species and in case of lacking live specimens. Henley et al. (2006) described integumental swabbing as a noninvasive method for cell collection from freshwater mussels. A more recent study presented a method for DNA extraction from empty apple snail shells (Andree and Lopez, 2013).

To avoid the phenomenon phenotypic plasticity and facilitate genetic analysis through small shell fragments from live or empty shells, we examined genetic diversity in four populations of freshwater mussels using 28S rDNA sequences. Here we present a successful method and first genetic data on Turkish populations of freshwater mussels.

Materials and Methods

Live specimens and empty freshwater mussel shells

Freshwater mussels and empty shells were collected from Gölarmara/Manisa (lake), Manavgat/Antalya (dam lake downstream), Küçük Akgöl/Adapazarı (lake), Karasu/Sinop (river) in Turkey (*Table 1*) with the permission of the Ministry of Agriculture and Forestry (*Figure 1a*). While empty shells were placed in plastic bags, tissue samples were taken from live specimens into ethanol 70% and transferred to the laboratory in cooler boxes. Samples were stored at -20°C until analysis.

Table 1. Sampling locations

Location	Waterbody	Coordinates	Number of specimens
Karasu	Stream	41.56162"N – 35.04827"E	150
Küçük Akgöl	Lake	40.52365"N – 30.25582"E	40
Gölarmara	Lake	38.35449"N – 27.59487"E	4
Manavgat	Dam Lake	36.52434"N – 31.31211"E	11

DNA preparation, PCR and sequencing

Total DNA was extracted using EurX GeneMATRIX Tissue & Bacterial DNA Purification Kit (Poland) according to the manufacturer protocol with minor modifications. Samples from each shell were removed with an 8 mmØ ceramic cutting hand drill and crushed into fine powder in a mortar. To avoid contamination, cutting edges were changed with new between samples. Sample material was placed into the 2 ml tube and 350 µl Lyse T buffer added and the precipitate was suspended. Tissue samples were cut into small pieces and buffered. 2 µl RNaseA and 20 µl Proteinase K were added and vortexed. Samples were incubated at 56°C during 3 h after adding 350 µl buffer, the samples were incubated for 10 min at 70°C. Pure ethanol of 350 µl was added and the tubes were centrifuged for 1 min at 12 000g. The supernatant was transferred into DNA binding spin-column and centrifuged repeatedly. After placing the column in the collection tube 500 µl washing buffers were added and centrifuged. Elution buffer was added in new tubes and incubated at room temperature. DNA products were stored at 4°C. The recovered DNA was checked with Thermo Scientific Nanodrop 2000 (USA) for integrity.

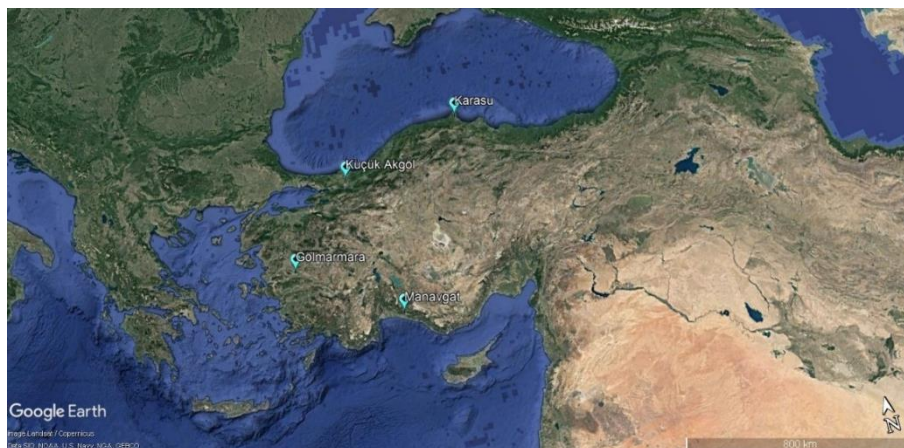


Figure 1a. Map showing the sampling locations

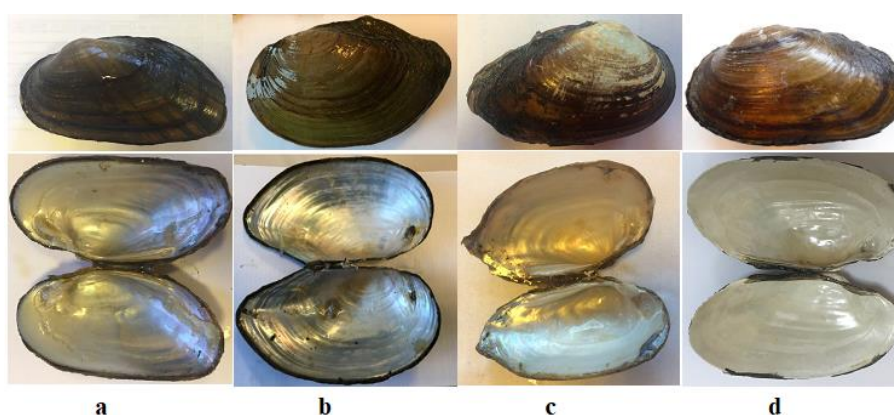


Figure 1b. Shells from collection sites. (a)Karasu, (b)Akgöl, (c) Göl marmara and (d) Manavgat

A universal primer set (Justine et al., 2002) was employed for mitochondrial 28S rDNA gene amplification (Table 2).

Table 2. Primer set used for 28S rDNA amplification

Primer	Direction	Primer Sequence	Reference
C1	Forward	5'-ACCCGCTGAATTTAAGCAT-3'	Justin et al., 2002
C2	Reverse	5'-CTCTCTYTYCAAAGTTCTTTTC-3'	

Each PCR reaction mixture contained 1.5 µl DNA template (880 ng/µl), 1 µl DNA FIREPol® DNA Polymerase (5 U/µl), 10 µl 10X Buffer, 1 µl dNTP (20 mM), 6 µl 25 mM MgCl₂ (1.5 mM) and 10 pmol/µl of each primer (0.15 µM), completed with ddH₂O to a total volume of 100 µl. The thermal cycler profile was as one cycle initial denaturation step at 95°C for 5 min, 35 cycles of denaturation at 95°C for 45 sec, annealing at 57°C for 45 sec, elongation at 72°C for 1 min and a final extension step at 72°C for 5 min. The amplified DNA fragments were controlled with agarose gel electrophoresis and purified using ExoSAP-IT™ PCR Product Cleanup Reagent"

(ThermoFisher Scientific, USA). PCR amplified DNA fragments were sequenced with automated 3500xL Genetic Analyzer (Applied Biosystems).

The sequences were aligned with Sequencer v5.4.6 and consensus alignments for each sampling location were deposited to GenBank under accession numbers MK530498, MK530499, MK530500, MK530501.

The number of polymorphic sites, DNA polymorphism, conserved regions, haplotype number, haplotype diversity and Tajima's D statistics were performed with DnaSP v.5.0 (Librado and Rozas, 2009).

Phylogenetic analysis for the 28S rDNA sequence set was performed using the software MEGA X (Kumar et al., 2018). The tree was produced with the Neighbor-Joining (NJ) method using the Kimura 2-parameter model for genetic distance estimation (Kimura, 1980). Percentage of replicate trees in the bootstrap test (1000 replicates) indicated alongside the branch nodes (Felsenstein, 1985).

Results

DNA extraction from all three types of material was affordable. *Figure 2* demonstrates the gel electrophoresis screenshot for mitochondrial DNA from mantle clipping, fresh shell and old empty shell material. Even though mitochondrial DNA can be isolated from all samples, sequencing performance was not the same. While 97% of tissue and 70% of fresh shell samples were sequenced successfully, only 10% of the PCR products from old empty shells served quality sequences (*Table 3*).

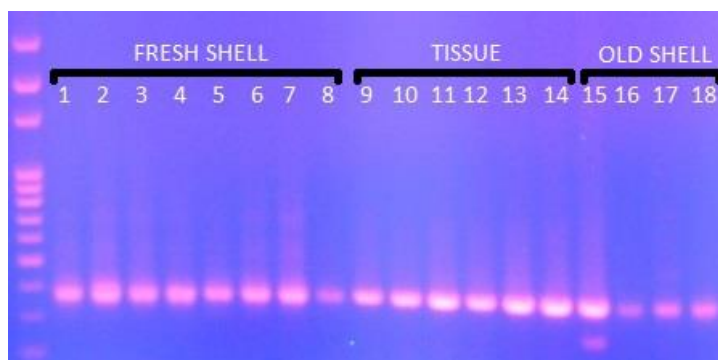


Figure 2. Gel electrophoresis of mantle tissue, fresh shell and old shell samples

Table 3. Number of successfully sequenced specimens from different DNA sources

Location	Tissue		Fresh Shell Material		Old Empty Shell Material	
	Template DNA (n)	Sequenced (n)	Template DNA (n)	Sequenced (n)	Template DNA (n)	Sequenced (n)
Karasu	10	10	10	10	10	2
Küçük Akgöl	10	10	10	7	10	-
Gölmarmara	6	5	6	4	10	1
Manavgat	10	10	10	4	10	1

Twenty-five 28S rDNA sequences have an average length of 339 bp. Mean G+C ratio found 54.1% ranging between a high scale of min 52.9% and max 54.8%. (*Table 4*). Specimens from each sampling location were given in consensus sequences

in NJ tree construction. Nucleotide compositions of consensus sequences are given in *Table 4*.

For all data set, 136 variable sites exist in 339 bp sequence (5.04%). *Table 5* shows conserved, variable and parsimony informative sites, identical, transitional and transversional pair frequencies and transition/transversion ratio (R). The data set has 82 polymorphic and 52 parsimony informative sites and transitional/transversional pair ratio of 0.8 (R).

Table 4. Nucleotide composition

Location	T (U)	C	A	G	G+C	A+T
Karasu	20.4	24.3	26.7	28.5	52.9	47.1
Küçük Akgöl	20.1	24.7	25.6	29.6	54.3	45.7
Gölmarmara	19.6	25.2	25.8	29.5	54.7	45.3
Manavgat	20.4	24.5	24.8	30.3	54.8	45.2
Avg.	20.1	24.7	25.7	29.5	54.1	45.9

Table 5. Nucleotide pair frequency

Conserved sites	199/339
Variable sites	136/339
Parsimony informative sites	52/339
Singleton sites	83/339
Identical pairs ii	259
Transitional pairs si	10
Transversional pairs sv	13
R si/sv	0.8

The sequence data set represented 16 haplotypes. While haplotype H9 existed in six samples (2 from Gölmarmara and 4 from Akgöl), haplotype H10 was found in three samples (1 from Gölmarmara, 1 from Akgöl and 1 from Manavgat). Other 14 haplotypes were found for each sample from all locations. Mean haplotype diversity (Hd), nucleotide diversity (π), Tajima's D test results are given in *Table 6*. Haplotype diversity changed between 0.714 and 1.000 min/max from Akgöl and Manavgat. Average Hd and π were calculated as 0.0881 and 0.062 respectively. Test of neutrality on allele frequency distribution with Tajima's D indices resulted negative for all four populations and was not statistically significant ($p>0.05$).

Table 6. Genetic diversity and neutrality results (Hd: Haplotype diversity, π : nucleotide diversity)

Location	n	h	Hd	π	Tajima's D
Karasu	10	9	0.977	0.065	-0.419
Küçük Akgöl	7	4	0.714	0.117	-1.543
Gölmarmara	4	3	0.833	0.006	-0.780
Manavgat	4	4	1.000	0.059	-0.635

Nucleotide substitution rates were calculated with maximum likelihood method using Tamura 3-parameter model (*Table 7*). The nucleotide frequencies are A=23.13%, T/U=23.13%, C=26.87%, and G=26.87%.

The divergence between four populations ranged between 0.019 and 0.045 with an average value of 0.031 (3.1%) indicating a correlation between divergence and location. While the highest diversity was found between Karasu and Manavgat populations (4.5%), the populations from Gölarmara and Akgöl were the genetically closest ones (1.9%) (*Table 8*). The high distance value of Karasu samples is due to the species difference. The Karasu population is clearly located on a separate branch in the phylogenetic tree given in *Figure 3*.

Table 7. Nucleotide substitution matrix

	A	T	C	G
A	-	6.77	7.86	11.15
T	6.77	-	11.15	7.86
C	6.77	9.60	-	7.86
G	9.60	6.77	7.86	-

Each entry is the probability of substitution (*r*) from one base (row) to another base (column). Rates of different transitional substitutions are shown in **bold** and those of transversional substitutions are shown in *italics*

Table 8. Average *p* distances for 28S rDNA between different geographic populations

n		1	2	3	4
4	Manavgat	-			
7	Akgol	0.028	-		
4	Golarmara	0.019	0.003	-	
10	Karasu	0.045	0.044	0.045	-

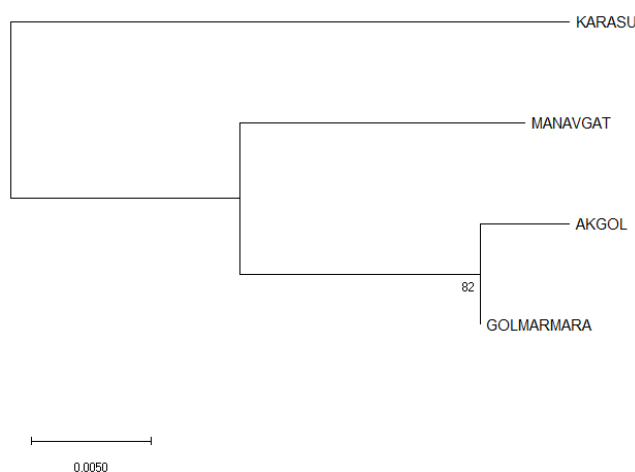


Figure 3. Phylogenetic relationships among four taxa inferred from 28S rDNA sequences. The evolutionary history was inferred using the Neighbor-Joining method (Saitou and Nei, 1987). The percentage of replicate trees in which the associated taxa clustered together in the bootstrap test (1000 replicates) are shown next to the branches (Felsenstein, 1985). The evolutionary distances were computed using the Kimura 2-parameter method (Kimura, 1980). Evolutionary analyses were conducted in MEGA X (Kumar et al., 2018)

Figure 3 shows the evolutionary relationship between the four populations. The tree was obtained by Neighbor-Joining method (Saitou and Nei, 1987) in MEGA X (Kumar

et al., 2018). The optimal tree with the sum of branch length was 0,058775. Genetic distances between four populations were estimated by Kimura 2-parameter method (Kimura, 1980). The percentage of replicate trees in which the associated taxa clustered together in the bootstrap test (1000 replicates) are shown next to the branches (Felsenstein, 1985). The analysis involved 4 nucleotide sequences. All ambiguous positions were removed for each sequence pair. There were a total of 339 positions in the final dataset.

Discussion

This is the first study investigating the genetic diversity of freshwater bivalvia through shell DNA analysis. The results clearly indicated that 28S rRNA sequences obtained from shell material are informative to reveal major clades, haplotypes and phylogenetic relationship between four populations. These findings support that quite fresh shell material from freshwater mussels is to be a convenient DNA source for molecular-based studies. Degradation of shells due to the exposure time to environmental components, basically the water, is the primary limiting factor for shell DNA.

G+C ratio (average 54.1%) represented in *Table 4* indicates heterogeneity between four populations. The high number of variable sites supports that the 339 bp 28S rRNA sequence from the shell is a successful molecular marker for freshwater mussels. Transitional/Transversional pair rate R (si/sv) found 0.8 while substitution rate R which is calculated using Kimura-2 parameter was 0.7. These close rates also support the reliability of data analysis.

Consensus sequence analyses supported two well-differentiated clades of unionid species swan mussel *Anodonta cygnea* from Manavgat, Akgol and Golmarmara Lakes and thick shelled river mussel *Unio crassus* from Karasu River. Genetic distance between populations found between 0.003 and 0.045 (*Table 8*) with an average of 0.030. Akgol and Golmarmara populations have exhibited the lowest genetic distance while Karasu River samples have a constant high distance value which is comparable with species difference.

According to Gürlek et al. (2014) Modell has defined four main groups of freshwater mussels of Anatolia according to geographic distribution containing Unio, Anodonta, Potomida, Leguminaia, Gabillotia, Pseudodontopsis, Sinanodonta species. There are only a few more studies on freshwater mollusks of Turkey which all used the morphological description of species. This study is being the first as using shell DNA for genetic classification of freshwater bivalvia from Turkey. Due to the difference between sampling locations, there is a lack of chance to conclude data from the literature. Only Şahin (2013) defined bivalvia specimens from Karasu as *Unio pictorum* through morphological identification methods but our findings indicated a different taxonomy of specimens from the same locality as *Unio crassus* according mtDNA sequence.

The phylogenetic tree of 28S rRNA sequences clearly revealed the genetic difference between four populations (*Figure 3*). Karasu samples were located on a separate branch. Akgol and Golmarmara populations stand close with a result of the hypotheses of geographic isolation.

Ricciardi and Rasmussen (1999) have reported the rapid extinction of freshwater species in North America and made an extinction rate estimation for freshwater mussels

about two fold than the average of freshwater species. Lopes-Lima et al. (2017) have underlined the lack of data on freshwater mussels especially in south-eastern European countries and the importance of more studies for effective management measures.

While thick shelled river mussel *Unio crassus* is classified as “globally endangered” on IUCN Red List (Lopes-Lima et al., 2014), the swan mussel *Anodonta cygnea* is referred as “near threatened” by Killeen and Aldridge (2011). Also, it is noted that the population trend of *Unio crassus* is unknown for the Mediterranean basin so referred to DD (Data Deficient) by Vavrova (2011). Since the taxonomy of freshwater bivalve is still ambiguous and species distribution and population sizes are unclear, molecular studies can serve valuable information for conservation management plans.

Conclusion

The need for effective conservation and management strategies for freshwater ecosystems which are under intense pressure, especially human activities, requires wider monitoring using standardized protocols and methods. But there are gaps on population structures, distribution and taxonomic data for many species as well as freshwater mussels. On country basis, large scale taxonomic studies which are supported with molecular data will help filling these gaps. International cooperation is strongly suggested for database and management projects.

Overall this study suggests shell DNA as a reliable source of divergence data on freshwater bivalve species against the problematic systematic relationships and phenotypic plasticity. Partial sequences of 28S rDNA derived from four populations provide the first genetic data in Turkey with this study.

Future researches are needed for large scale data collection on freshwater bivalves by using non-invasive standardized methods.

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REFERENCES

- [1] Andree, K. B., López, M. A. (2013): Species identification from archived snail shells via genetic analysis: a method for DNA extraction from empty shells. – *Molluscan Research* 33(1): 1-5.
- [2] Baker, A. M., Sheldon, F., Somerville, J., Walker, K. F., Hughes, J. M. (2004): Mitochondrial DNA phylogenetic structuring suggests similarity between two morphologically plastic genera of Australian freshwater mussels (Unionoida: Hyriidae). – *Molecular Phylogenetics and Evolution* 32(3): 902-912.
- [3] Bogan, A. E. (2008): Global diversity of freshwater mussels (Mollusca, Bivalvia) in freshwater. – *Hydrobiologia* 595: 139-147.
- [4] Felsenstein, J. (1985): Confidence limits on phylogenies: An approach using the bootstrap. – *Evolution* 39: 783-791.
- [5] Gürlek, M., Kara, C., Kebapçı, Ü. (2014): Adıyaman Azaplı Gölü'nde yaşayan *Unio terminalis* (Bourguignat, 1852)'in bazı konkolojik özellikleri. – *Aquaculture Studies* 2014(3): 23-28.
- [6] Henley, W. F., Grobler, P. J., Neves, R. J. (2006): Non-invasive method to obtain DNA from freshwater mussels (Bivalvia: Unionidae). – *Journal of Shellfish Research* 25(3): 975-978.

- [7] Justine, J. L., Jovelin, R., Neifar, L., Mollaret, I., Lim, L. H. S., Hendrix, S. S., Euzet, L. (2002): Phylogenetic positions of the Bothitrematidae and Neocalceostomatidae (monopisthocotylean monogeneans) inferred from 28S rDNA sequences. – *Comparative Parasitology* 69(1): 20-25.
- [8] Killeen, I., Aldridge, D. (2011): *Anodonta cygnea*. – The IUCN Red List of Threatened Species 2011: e.T156066A4907255. Downloaded on 25 April 2019.
- [9] Kimura, M. (1980): A simple method for estimating evolutionary rate of base substitutions through comparative studies of nucleotide sequences. – *Journal of Molecular Evolution* 16: 111-120.
- [10] Kumar, S., Stecher, G., Li, M., Knyaz, C., Tamura, K. (2018): MEGA X: Molecular Evolutionary Genetics Analysis across computing platforms. – *Molecular Biology and Evolution* 35: 1547-1549.
- [11] Librado, P., Rozas, J. (2009): DnaSP v5: A software for comprehensive analysis of DNA polymorphism data. – *Bioinformatics* 25: 1451-1452.
- [12] Lopes-Lima, M., Kebapçı, U., Van Damme, D. (2014): *Unio crassus*. – The IUCN Red List of Threatened Species 2014: e.T22736A42465628. <http://dx.doi.org/10.2305/IUCN.UK.2014-1.RLTS.T22736A42465628.en>. Downloaded on 25 April 2019.
- [13] Lopes-Lima, M., Sousa, R., Geist, J., Aldridge, D. C., Araujo, R., Bergengren, J., Douda, K. (2017): Conservation status of freshwater mussels in Europe: state of the art and future challenges. – *Biological Reviews* 92(1): 572-607.
- [14] Ricciardi, A., Rasmussen, J. B. (1999): Extinction rates of North American freshwater fauna. – *Conservation Biology* 13: 1220-1222.
- [15] Saitou, N., Nei, M. (1987): The neighbor-joining method: A new method for reconstructing phylogenetic trees. – *Molecular Biology and Evolution* 4: 406-425.
- [16] Strayer, D. L. (2008): *Freshwater Mussel Ecology: A multifactor approach to distribution and abundance*. – University of California Press, Berkeley, CA.
- [17] Şahin, S. K. (2013): Mollusca species of Down Sakarya River (Karasu) and some physico-chemical variables effecting their abundance. – *Aquaculture Studies* 13: 011-019.
- [18] Vavrova, L. (2011): *Unio crassus*. – The IUCN Red List of Threatened Species 2011: e.T22736A9382322. Downloaded on 25 April 2019.